

*British Journal of Environment & Climate Change 6(4): 259-278, 2016, Article no.BJECC.2016.025 ISSN: 2231–4784*



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# **Hydrobiological Assessment of the High Mountain Kolsay Lakes (Kungey Alatau, Southeastern Kazakhstan) Ecosystems in Climatic Gradient**

**E. G. Krupa1 , S. M. Barinova2\*, S. M. Romanova3 and A. B. Malybekov4**

*1 Institute of Zoology, Ministry of Education and Science, Almaty 050060, Kazakhstan. <sup>2</sup> Institute of Evolution, University of Haifa, Mount Carmel, 199 Abba Khoushi Ave., Haifa 3498838, Israel. <sup>3</sup> Al-Farabi Kazakh National University, Ministry of Education and Science, Almaty 050000,* 

*Kazakhstan. <sup>4</sup> National Park "Kolsay Lakes", Small Settlement Saty, Ultarakova Street, 38, Almaty Region,* 

*Kazakhstan.*

# *Authors' contributions*

*This work was carried out in collaboration between all authors. Author EGK designed the study, wrote the protocol and wrote the first draft of the manuscript. Author SMB performed the statistical and bioindicational analysis. Authors SMR and ABM managed the analyses for the study. All authors read and approved the final manuscript.*

# *Article Information*

DOI: 10.9734/BJECC/2016/26496

*Original Research Article*

*Received 27th March 2016 Accepted 8th June 2016 Published 2nd December 2016*

# **ABSTRACT**

**Aims:** The aim of this study is to assess the ecological status of the Kolsay high mountain lakes according to their hydrochemical parameters and structure of phytoplankton communities and reveal the altitude climatic condition impact.

**Study Design:** We implemented bio-indication and diverse statistical methods, which represent some new approaches in freshwater algal diversity analysis.

**Place and Duration of Study:** Institute of Evolution, University of Haifa, Israel, Institute of Zoology, Ministry of Education and Science, Almaty, Kazakhstan, between August 2015 and April 2016.

**Methodology:** We collected 20 samples of phytoplankton in August 2015 from four ultrafresh high mountain Kolsay lakes, located at the altitude of 1829-3170 meters above sea level in the Kungey Alatau, Southeastern Kazakhstan in gradients of climatic and environmental variables that we analyzed. Bio-indication and statistical methods for the ABC Method, Shannon-Weaver index, and WESI index (Water Ecosystem State Index) that indicated some toxic effect on photosynthesis of

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algae were used together with Statistica 12.0 and CANOCO Programs for calculating of relationships between environmental and algal communities data among the altitude of the lakes. **Results:** As a result, 28 species from five taxonomic divisions were identified in phytoplankton communities. According to hydrochemical indices, the lakes were classified as clean, with a slightly increased level of organic pollution in the Middle and Upper Kolsay. The ratio of species in the phytoplankton communities in all the lakes was corresponding to alkaline fresh non-flowing waters. Indication of the indicator species' ratio, functional groups and size structure (the Abundance-Biomass-Comparison method, ABC) of phytoplankton showed that the level of organic pollution decreased in the high-altitude direction. The relationship between the graphic (ABC Method) and the calculation method (Shannon-Weaver index) used for the analysis of the phytoplankton structure is shown. Dynamics of saprobiological and dimensional parameters of phytoplankton links well with the distribution of the recreational load on the lake. The change of biomass of planktonic algae to a greater extent was determined by the dynamics of nitrogen compounds. In general, together with a low level of heavy metals in the water of lakes, the values of WESI index (Water Ecosystem State Index) indicated some toxic effect on photosynthesis of algae in the Lower, Middle Kolsay and the Lake Sary-Bulak. We revealed major variables that considered stimulating or stress factors with helps of statistical programs.

**Conclusion:** Analysis of hydrochemical and biological data demonstrated that Kolsay mountain lakes are in the early stages of eutrophication. The processes of eutrophication are most evident in the Lower Kolsay. The changes in ecosystems of Kolsay mountain lakes can be caused by the growth of recreational load alongside the altitude related climate change.

*Keywords: Mountain lakes; nutrients; heavy metals; phytoplankton; bio-indication; the ABC-method; the Shannon-Weaver index.*

### **1. INTRODUCTION**

The water quality management on the watersheds is the goal of modern management, which is impossible without an understanding of the structural changes in the biotic community and identifying the main factors that determine the water quality of aquatic object [1]. Regarding the procedures for assessing the ecological status of natural ecosystems, more and more attention is paid to single-celled algae [2-4]. In order to assess the level of organic pollution of water bodies, different structural indicators of phytoplankton communities are used: the ratio of species  $-$  the saprobity indicators [5], the biomass [6], a variety of functional groups and the ratio of diversity to biomass [7,8], the ratio of taxonomic divisions [9,10], size classes of diatoms [11], the value of the average mass of algal cell [12,13].

The graphs of rank distributions of species serve as an integral reflection of community structure [14]. Abundance-Biomass-Comparison (the ABC-Method) has been applied to analyze the structure of benthic [15-17] and planktonic communities [18-21] in the gradient of external factors.

The high mountains Kolsay lakes are located in the Kungey Alatau Mountains [South-Eastern Kazakhstan] within the territory of the National Natural Park, established in 2007. The lakes are remote from the agricultural and industrial areas. Anthropogenic impact on their ecosystems is associated with recreational load and acclimatization activities. Due to the inaccessibility, the Kolsay Lakes are poorly explored. Fragmentary information is only available on zooplankton and hydrochemical parameters of the two lower lakes [21-23].

Mountain cold-water lakes are more sensitive to external factors compared to lowland reservoirs. Significant changes in hydrocenosis of mountain lakes can occur even in the absence of direct human impact [24]. In order to preserve the mountain lakes as a specially protected object, the methods for capturing the changes on the early stages are needed. The aim of this study is to assess the ecological status of the Kolsay mountain lakes according to their hydrochemical parameters and structure of phytoplankton communities.

#### **2. MATERIALS AND METHODS**

Complex studies of the Lower, Middle, and Upper Kolsay Lakes and Sary-Bulak Lake (Fig. 1) were carried out in August 2015. In the field environment, the temperature, pH and electrical conductivity of the surface water layers were measured by means of HANNA HI 98129 devices. Water transparency was measured by Secchi disk. On the each lake, a sample of 1 liter of phytoplankton was taken. Each sample consisted of subprobe selected in three to five different parts of the lake. Subprobes were mixed, and then one integrated sample of required volume was selected. The water samples to determine the mineralization and chemical composition, the content of easily oxidized organic matter, nutrients and heavy metals were collected using the same principle.

The conventional methods of chemical analysis of water were used [25,26]. All water samples were analyzed in three-four replications. Determination of heavy metals in water was performed by mass-spectrographic method with inductively coupled plasma using the Agilent 7500A manufactured by the Agilent Technologies, USA (ST RK ISO 17294-2-2006). Assessment of the ecological state of the lakes on the hydrochemical parameters were carried out according to [6]. Data on the content of nutrients and heavy metals were compared to

the maximum permissible concentrations of substances for fishery waters (MPC) [27,28].

For the processing of phytoplankton samples, the settling method was used [29]. Species identification of planktonic algae was performed by using determinants for relevant divisions [30- 35]. Shannon-Weaver index was calculated on Primer 5 program through the logarithm with the base 2 in two versions – on the share in the total number of species [bit/individual] and on biomass [bit/mg] [36]. Curves on domination were calculated using the same program as the Abundance-Biomass-Comparison method (ABC) [14] which we implemented earlier for phytoplankton [18-21]. W-statistic of Clarke was calculated automatically [37]. Its value indicates the position of the biomass in respect to the curve the abundance. A positive W value indicates that the biomass curve is higher than abundance curve, and vice versa. The average mass of the algal cell was found as the ratio between the total biomass and the total abundance of phytoplankton.



**Fig. 1. Map – location scheme of the Kolsay lakes**

The ecological characteristics of algal species were obtained from the database compiled for freshwater algae of the world from multiple analyses of algal biodiversity by S.S. Barinova et al. [3], according to substrate preference, temperature, oxygenation, pH, salinity, organic enrichments, N-uptake metabolism, and trophic states. The ecological groups were separately assessed according to their significance for bioindications. Species that respond predictably to environmental conditions were used as bioindicators for particular variables of aquatic ecosystems, the dynamics of which are related to environmental changes. The statistical methods used were those recommended by Heywood [38] for the development of taxonomic studies, namely the CANOCO program [39] for canonical correspondence analysis (CCA) [40,41], the GRAPHS program [42] for Principal Component Analysis (PCA) and comparative floristics. Surface plots for biological and environmental variables relationship analysis were constructed by using the Statistics 12 program.

The correspondence between the classificational indicators of ecosystem according to the biogenic elements and those of saprobity indices was conducted through calculating the index of the ecosystem state (WESI, Water Ecosystem State Index) as the quotient from the classification rank on biota divided by the classification rank on the environment. The WESI index ranges from 0 to 9. If it is less than one, then the ecosystem is exposed to the toxic effects; if equal to or greater than  $1 -$  the self-purification is not suppressed [3,43].

# **3. RESULTS**

# **3.1 General Characteristics of the Study Region**

The Kolsay lakes are located in the Kungey Alatau Mountains (South-Eastern Kazakhstan) at the altitude from 1829 up to 3170 m a.s.l. Three lakes (Lower, Middle and Upper Kolsay) are flow through, canyon type, deep-water, small in

surface area, with high water transparency (Table 1). Originating in the northern and western slopes of the mountains, the river of the same name flows through the lakes. Two more small rivers fall into the Middle Kolsay. The Sary-Bulak Lake is shallow and located in the alpine zone. It is supplied by groundwater and precipitation. A small stream is flowing outflows from the lake and flows into the Kolsay River. Water transparency is to the bottom. The bottom of the lake is overgrown with Charophyta algae.

Lower, Middle and Upper Kolsay lakes are coldwater. The temperature of water decreases with altitude. The water temperature in Sary-Bulak Lake is higher than in the Middle and Upper Kolsay. In August 2015, immediately after the snowfall, its water temperature reached 12.7°C. In 2002 and 2006, the water was heated to a temperature  $18-20^{\circ}$ C [21]. The water in all lakes is alkaline, with a maximum pH in the Sary-Bulak. Lakes are ultrafresh (Table 2). Mineralization of water decreases in the highaltitude direction. Water is of the carbonate class of calcium group, the second type, very soft. There was small amount of readily oxidized organic matter and nutrients in the water of all lakes. The amount of nitrite in the water decreased in the direction from the lower to the upper lake. Nitrate content varied within a relatively small range. In the Sary-Bulak Lake, the nitrates were not recognized. Ammonium ions were absent in the water of the Lower Kolsay, and in other lakes the concentration changed insignificantly. Phosphorus was not in water.

The content of heavy metals in water, with the exception of copper, were one or two orders lower than the established permissible concentrations of substances [27,28]. The concentration of lead, copper, zinc, nickel and chromium in water decreased from Lower to Upper Kolsay. In the Sary-Bulak Lake, the content of almost all heavy metals were higher than in the downstream lakes.

**Table 1. Hydro physical characteristics of Kolsay Lakes, August 2015**



Variable	<b>Units</b>	Lower Kolsay Middle Kolsay		<b>Upper Kolsay</b>	Kolsay	
					Sary-Bulak	
Hardness	mg Eq. $d\overline{m^3}$	1.4	1.2	1.1	0.25	
Deep	m	36.6	54.0	25.0	2.5	
Secchi	m	9.0	9.0	9.0	2.5	
TDS	mg dm $3$	123.9	116.0	102.2	26.6	
<b>BOD</b>	mgO dm <sup>-3</sup>	1.36	1.2	1.2	2.14	
NO <sub>2</sub>	mg $dm^{-3}$	0.028	0.03	0.02	0.011	
$N-NO2$	mg dm $3$	0.008	0.009	0.006	0.003	
NO <sub>3</sub>	mg dm $^{-3}$	0.688	0.344	0.516	0.0	
$N-NO3$	mg $dm^{-3}$	0.155	0.078	0.117	0.0	
NH <sub>4</sub>	mg $dm^{-3}$	0.0	0.2	0.15	0.22	
$N-NH_4$	mg $dm^{-3}$	0.0	0.156	0.117	0.171	
PO <sub>4</sub>	mg dm $3$	0.0	0.0	0.0	0.0	
Fe	mg dm $3$	0.140	0.164	0.044	0.440	
Si	mg dm $3$	4.2	6.3	4.4	1.0	
Mn	mg dm $^{-3}$	0.0165	0.008	0.008	0.010	
Pb	mg dm $3$	0.00014	0.00009	0.0001	0.0005	
Cu	mg $dm^{-3}$	0.0038	0.0031	0.0026	0.0055	
Zn	mg dm $^{-3}$	0.0009	0.0010	0.0002	0.0016	
Cd	mg dm $3$	0.00045	0.0004	0.0006	0.0004	
Ni	mg dm $3$	0.0018	0.0017	0.0014	0.0027	
Cr	mg $dm^{-3}$	0.0008	0.0007	0.00065	0.0005	
No of Species		15	8	10	6	
Cyanobacteria		3	0	0	0	
Bacillariophyta		$\overline{7}$	8	8	1	
Charophyta		$\pmb{0}$	0	0	1	
Chlorophyta		3	0	$\overline{2}$	4	
Euglenophyta		$\overline{2}$	0	0	0	
Abundance	mln. cells $m-3$	510.2	75.0	110.1	11.7	
Biovolume	mg $m-3$	225.7	450.7	477.3	78.3	
Shannon Ab	bit/individual	2.16	2.38	2.67	2.24	
Shannon Bi	bit/ma	2.61	1.94	2.25	1.14	
Average cell mass	$×10^{-6}$ , mg	0.442	6.009	4.335	6.692	
Average W		$-0.077$	0.127	0.136	0.610	
Index Saprobity S		1.51	1.12	0.99	0.76	
Index WESI		0.50	0.67	1.00	0.50	

**Table 2. Chemical and biological variables in the Kolsay lakes, August 2015**

#### **3.2 Brief Description of the Lakes Phytoplankton**

Altogether 28 species of phytoplankton were identified in Kolsay lakes (Table 3). The major number of them was Bacillariophyta. The highest algal diversity in Division level was recorded in the Middle and Upper Kolsay. Bluegreen algae were found only in the Lower Kolsay.

Phytoplankton in Lower and Middle Kolsay lakes was close in species composition (Fig. 2a). The algae communities of Upper Kolsay and Sary-Bulak were distinguished into separate groups. Similar results were obtained when using PCA analysis (Fig. 2b).

The number of phytoplankton cells varied by orders of values, with the maximum index value in the Lower Kolsay and its tendency to decrease with altitude (Table 2). In the Lower Kolsay Lake, blue-green algae dominated up to 90% of the total abundance. Dominant complex included *Planktolyngbya contorta* (20.6%), *P. limnetica*  (36.3%), and *Geitlerinema amphibium* (33.3%). Only diatoms were represented in the Middle<br>Kolsay phytoplankton. Diatoms (69.7%) Kolsay phytoplankton. Diatoms (69.7%) dominated numerically, and green algae (30.3%) sub-dominated in the Upper Kolsay. In these two lakes, *Lindavia comta, C. meneghiniana, C. planctonica* made the largest contribution to the formation of the community composition (9.1–37.7%). The dominant species complex in the Upper Kolsay also included *Sphaerocystis planctonica* (24.3%) from the green algae. Green algae formed up to 85.5% and diatoms up to 14.5% of abundance in phytoplankton of the Sary-Bulak Lake. The most numerous were diatoms *Lindavia comta* (14.5%), green algae *Closteriopsis longissima* (14.5%), *Monoraphidium contortum* (28.2%), *M. obtusum*  (14.5%), as well as *Spirogyra* sp. (28.2%) from Charophyta Division.

The phytoplankton biomass was highest in the Middle and Upper Kolsay lakes. The biomass in the Sary-Bulak Lake was shaped with 69.7– 100.0% by diatoms. The euglenoids were subdominants in the Lower Kolsay Lake making up to 22.9% of the phytoplankton biomass. The Sary-Bulak Lake phytoplankton biomass was represented by green algae, which made the main contribution (81.6%), with diatoms (18.4%) as sub-dominants. The diatoms *Cyclotella meneghiniana* (20.0–35.4%) and *C. planctonica*  (27.2–38.6 %) occupied the dominant position on biomass in the three lower lakes. The dominant complex in Lower Kolsay was included *Trachelomonas intermedia* (15.5%). *Lindavia comta* played a significant role in the formation of phytoplankton biomass of Middle, Upper Kolsay and Sary-Bulak (18.4–34.4%) lakes. The filamentous algae *Spirogyra* sp. made the largest contribution (74.5%) to the formation of the Sary-Bulak Lake phytoplankton biomass.

Shannon-Weaver index was characterized by the moderate level of the diversity of the algal communities of the lakes (Table 2). The phytoplankton average cell mass was enlarged with increasing altitude.

The species richness of phytoplankton communities grew up with the nitrates increase, but not with water mineralization (Fig. 3a), and correlated with the increase in abundance (Fig. 3b) as well. The number of diatom species in phytoplankton grew up with increasing amounts of nitric nitrogen and was highest in smallspecies communities (Fig. 3c), subject to water silicon saturation (Fig. 3d).

# **3.3 Assessment of the Ecological State of the Lakes Ecosystem by Hydrophysical and Hydrochemical Indicators**

Concentrations of heavy metals in the Kolsay lakes water were insignificant, which is typical for the mountain reservoirs of South-Eastern Kazakhstan [48,49]. Most of the analyzed parameters of the environment did not exceed the limits for extremely pure and clean water (Ι–ΙΙ Class, Rank 1–3), except for nitrite [6] (Table 4). The value of nitrites corresponded to Water Class ΙΙΙ – satisfactory purity in Lower, Middle and Upper Kolsay lakes, and to the level of pure water (Class I) in the Sary-Bulak Lake.

All lakes have alkaline water (Table 2). According to the classification [6], such pH values are characteristic of moderately polluted (three lower lakes) and extremely polluted waters (Sary-Bulak Lake). Evidently, this classification is cannot be used for assessing the ecological status of the Kolsay lakes. High water alkalinity in most water bodies of Kazakhstan is natural and not related to anthropogenic pollution. The pH values higher than 8.4 were observed even in the spring waters of Kazakhstan [50].



**Fig. 2. Dendrite showing the similarities in composition of phytoplankton species of the Kolsay lakes (a) and Principal Component Analysis (РСА) (b), August 2015**



# **Table 3. Taxonomy and ecology of the algal communities in the Kolsay lakes**

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Taxa	Code	LK	МK	UK	KS	Hab		Reo	рH	Hal	D	s	Sap	Tro	Aut- Het
Lemmermann															
Crucigenia quadrata Morren	CruQua	0			0	$P-B$	$\overline{\phantom{a}}$	st-str	acf			1.9	b-o		
Monoraphidium contortum (Thuret) Komárková-	MonCon	0	0		1	P	$\overline{\phantom{0}}$	$\overline{\phantom{0}}$	-						$\overline{\phantom{0}}$
Legnerová															
Monoraphidium griffithii (Berkeley) Komárková-	MonGri			$\Omega$	0	P						$2.2^{\circ}$	- b		
Legnerová															
Monoraphidium obtusum (Korshikov) Komárková-	MonObt	0	0			P						$2.2^{\circ}$	h.		
Legnerová															
Sphaerocystis planctonica (Korshikov) Bourrelly	SphPla	0			0	P									
Charophyta															
<i>Spirogyra</i> sp. $\sim$ $\sim$ $\sim$ $\cdots$ $\cdots$ $\cdots$ $\cdots$	Spirog	$\Omega$				B	$\overline{\phantom{0}}$					.		.	

*Note: LK – Lower Kolsay, MK – Middle Kolsay, UK – Upper Kolsay, KS – Kolsay Sary-Bulak. Note: Substrate preferences (Habitat): P – planktonic, P-B – plankto-benthic, B – benthic, S – soil. Temperature preferences (Temp): temp – temperate temperature, eterm – eurythermic, warm – warm-water. Oxygenation and streaming (Reo): st – standing water, str – streaming water, st-str – low streaming water. Halobity degree according Hustedt [44] (Hal): i – oligohalobes-indifferent, hl – halophiles. Acidity (pH) degree according Hustedt [45]: alf – alkaliphiles, ind – indifferents; acf – acidophiles. Organic pollution indicators according to Watanabe et al. [46] (D): sx – saproxenes, es – eurysaprobes, sp – saprophiles. Species-specific Index of Saprobity (S). Self-purification zone preferences (Sap): x – xenosaprob; o – oligosaprob; o-b – oligo-betamesosaprob; o-a – oligo-alpha-mesosaprob; b – beta-mesosaprob; b-a – beta-alpha-mesosaprob; a – alpha-mesosaprob. Nitrogen uptake metabolism (Aut-Het) [47]: ats – nitrogen-autotrophic taxa, tolerating very small concentrations of organically bound nitrogen; ate – nitrogen-autotrophic taxa, tolerating elevated concentrations of organically bound nitrogen; hne – facultatively nitrogen-heterotrophic taxa, needing periodically elevated concentrations of organically bound nitrogen. Trophic state (Tro) [47]: ot – oligotraphentic; o-m – oligo-mesotraphentic; m – mesotraphentic; me – meso-eutraphentic; e – eutraphentic*

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**Fig. 3. Surface plots for biological and environmental variables relationship analysis in the Kolsay lakes**





# **3.4 Assessment of the Ecological State of the Lakes by Phytoplankton**

Diatoms dominated in the phytoplankton communities in the three lower lakes, and green algae prevailed in the Sary-Bulak Lake (Fig. 4). Most of the phytoplankton species from the general taxonomic list belonged to alkaliphilic and alkalibiontes that prefer water with pH higher than 7.0, as well as to indifferents [3]. More than half of planktonic species indicators (56%) were represented by category indifferent to the water flow and oxygen saturation or prefer standing water bodies (Table 3). In relation to the mineralization, oligohalobes were dominant, among which the main part accounted for the indifferents.

The following changes occurred in the composition of phytoplankton with increase in altitude: the proportion of green algae species that live mainly in the plankton, the indicators of moderate temperatures, preferring standing water bodies and waters poorly enriched with oxygen, and alkaliphilic, increased; and the proportion of salinity indicators, with prevalence of oligohalobes – indifferents, decreased (Fig. 4).

Of the total taxonomic list of algae, saprobic valence is known for 21 species. The communities of all studied lakes were dominated by species – indicators of moderate pollution (Class ΙΙΙ of water quality) (Fig. 5). The composition of phytoplankton in the Lower and Middle Kolsay included an equal number of species that are typical for pure and polluted water. The latter were recorded in the Upper Kolsay. The species composition of phytoplankton became more homogeneous in relation to organic pollution in the high-altitude direction.

The dominant species make the maximum contribution to the assessment of the level of organic pollution. Blue-green algae *Planktolyngbya limnetica* and *Geitlerinema amphibium* (Lower Kolsay) are characteristic for clean and moderately polluted waters. *Lindavia comta* (Bacillariophyta) prefers waters with low levels of organic pollution (Middle, Upper Kolsay and Sary-Bulak), and *Cyclotella meneghiniana* is an indicator of organic pollution (Lower, Middle and Upper Kolsay). Green algae *Monoraphidium obtusum*, *M. contortum*, and *Closteriopsis longissima* (Sary-Bulak) are characteristic for clean and moderately contaminated water.

According to the average values of saprobity index (Table 2), water of the Lower Kolsay was assessed as moderately polluted (βmezosaprobic zone), Class III of water quality, and the Middle, Upper Kolsay and Sary-Bulak – as pure (oligosaprobic zone), Class II. Similar results were obtained when only diatoms were used for bioindication. According to Watanabe classification, the number of diatoms species, characteristic of pure water (saproxenes and eurysaprobes), increased, and the number of saprophiles decreased in the high-altitude direction (Fig. 5).

Proportion of the eutrophic water indicators increased with altitude, but indicators of trophic state as a system are only known for diatoms, so the Sary-Bulak Lake was not assessed (Fig. 5). Indication of nutrition type, is represented by diatoms only, showed that the proportion of strictly autotrophic species (ats) and species, enduring a low organic load (ate), decreased from upper to lower lakes and the proportion of facultative heterotrophic (hne) increased (Fig. 5). The WESI index varied from 0.50 to 0.67 in the three lakes, showing suppression of algal communities and only in the Upper Kolsay Lake it was equal 1.

Canonical correspondence analysis of the relationship between the environmental variables and dominant species distribution over the lakes communities (Fig. 6a), or over the taxonomic Divisions (Fig. 6b) carried out on the basis of Tables 1, 2 and 3, showed significant differences between the Sary-Bulak Lake, Lower Kolsay, and two intermediate, similar to each other, Middle and Upper Kolsay lakes. Fig. 6 shows that the environmental indicators stating the degree of influence on the algal communities divide into the following groups: trophic (arrows pointing down), affecting to a greater extend the phytoplankton of three lower lakes and stimulating the development of blue-green algae and euglenoids, and inorganic, associated with ground-water flow and affecting the community of the Sary-Bulak Lake.

According to the values of phytoplankton biomass (Table 2), the waters of the Lower, Middle and Upper Kolsay lakes were assessed as very clean (Class II, Rank 2), and in the Sary-Bulak as extremely pure (Class I, Rank 1) [6].

For further analysis, indicators for size structure of phytoplankton were drawn with W value statistics of Clarke, the value of average cell mass, and the value of the Shannon-Weaver index (bit/individual, and bit/mg).

Cumulative graphs of dominant species of planktonic algae showed changes in the size structure of communities in the high-altitude direction (Fig. 7). In Lower Kolsay, the biomass curve was above the abundance curve, i.e. Wstatistic of Clarke had a negative value. The W value became positive and increased in the

direction from the Middle Kolsay to upstream lakes.

The increase in the average cell mass (Table 2) also reflects the changes in the structure complexity of the phytoplankton in the highaltitude direction. In the Lower Kolsay, numerically dominant small-size blue-green algae occupied higher ranks (i.e. the proportion of the first dominant in the total abundance was higher than that of the first dominant in the total biomass) than larger diatoms, dominating in biomass (Table 5). Larger species of algae occupied higher ranks in biomass than in abundance for the phytoplankton of the three upstream lakes, respectively.







**Fig. 5. The ratio of the species number of species number - indicators of organic pollution, trophy and nutrition**  type in phytoplankton of the Kolsay lakes, August 2015. Abbreviation of ecological groups as<br>in Table 3 ators of organic pollution, trophy and nutrition<br>gust 2015. Abbreviation of ecological groups as<br>able 3<br>of the dominant species in the total abundance<br>of the Kolsay lakes, August 2015<br>ce, % Average cell Rank/ Biovolume, %









**(b)**

**Fig. 6. Canonical Correspondence Analysis of relationship between the variables of the Kolsay lakes environment and the dominant species (a), and Divisions (b) in phytoplankton. Abbreviation of species names as in Table 3**

Numeric expression of mutual position of dominance curves is the ratio of the Shannon-Weaver index values, calculated in two versions – bit/mg and bit/individual. The arithmetic difference between the values of the two versions was called ∆-Shannon-Weaver. There was a negative relationship between the amount

-1.0 1.5

 $-1.0$ 

rò  $\leftarrow$ 

> of W-value and values of ∆-Shannon-Weaver (Fig. 8).

#### **4. DISCUSSION**

Hydrochemical variables of the Kolsay lakes in August 2015 are corresponded to the levels of

extremely pure and clean water reservoirs (Ι – ΙΙ Class, Ranks 1–3), except for nitrites. According to the total points calculation (Table 3), the organic pollution decreased in the line "Middle Kolsay = Upper Kolsay > Lower Kolsay > Sary-Bulak. Phosphorus was not found in water of all lakes, which is characteristic of extremely pure water, according to hydrochemical classification [6]. Comparison with the data of previous years showed that in August 2002 the amount of phosphates in the water of the Lower and Middle Kolsay (Fig. 9) has reached the level of slightly polluted water. During the reported period of 2002, the total nitrogen content in the water of the Lower Kolsay Lake has increased from 0.143 to 0.716 mg dm $3$ , and in the Middle Kolsay, on the contrary, decreased from 1.286 to 0.574 mg dm<sup>-3</sup> [21].



**Fig. 7. The structure of species dominance in phytoplankton of the Kolsay lakes, August 2015**



**Fig. 8. Dynamics of the W-statistic (Clarke) and ∆-Shannon-Weaver**

Along with temperature, phosphorus and nitrogen are essential elements for controlling the development of planktonic algae [9,51-53]. In August 2015, in the absence of phosphorus in the waters of all lakes, phytoplankton biomass varied similarly to the dynamics of total nitrogen concentrations, with the highest values in the Middle and Upper Kolsay. It should be taken into consideration that the phosphate concentration at a given time is not an indicator of the productivity of the system, but only points out to the active phosphorus exhaustion by phytoplankton. That is, the low contents of mineral phosphorus may indicate that the system is either exhausted, or its metabolism is intense [54]. Obviously, this is also true for the Lower

Kolsay, where over the last decade there was a reduction of phosphates to analytical zero, and the concentration of total nitrogen increased almost twice, indicating the nitrogen/phosphorus system imbalance, with the latter being exhausted. The emergence of Charophyta algae in the coastal zone of the lake (Fig. 10) shows the intensification of in-water processes that were absent in 2002.

The use of biological indicators allowed capturing more subtle variations in the quality of the Kolsay lakes water. The ratio of species in phytoplankton communities of the Kolsay lakes as a whole matched the one for non-flowing alkaline waters. In terms of the ratio of indicator



**Fig. 9. The interannual dynamics of biogenic elements in the Kolsay lakes water**



**Fig. 10. The coastal area of the Lower Kolsay Lake in 2006 (left) and 2015 (right). Photo by Udartseva E.R. and Krupa E.G**

species and the ratio of functional groups of phytoplankton (type of nutrition), the level of organic pollution decreased in the high-altitude direction. According to the values of phytoplankton biomass, the water of the Lower, Middle and Upper Kolsay was assessed as very clean, and of the Sary-Bulak – as extremely clean.

In addition to the species composition, the size structure of communities has significant importance. In nutrient-rich conditions, the dominance of small-sized species (r-strategists) is a general principle for functioning of the biological systems [54]. The domination of smallsized species in planktonic [10-13,21,55,56] and benthic communities [15-17] enhances in the process of eutrophication of water bodies. Dimensional structure of diatoms is a sensitive indicator of even short-term changes in aquatic ecosystems [57].

The ABC-Method is an integral expression of community size structure. R. Warwick [14] showed that the biomass curves are above the abundance curves in the normal habitat of benthic organisms. With moderate stress, the curves are close to each other; with strong stress, the abundance curves are located above the biomass curves.

The values of W statistic of Clarke, showing the relative position of the curves of biomass and abundance, indicate a change in the structure of phytoplankton in the Kolsay lakes with reduction of stress in a high-altitude direction. The value of the indicator changed synchronically with the dynamics of the average mass of algal cell in community, and the values of ∆-Shannon-Weaver. Previously, different statistically significant negative relationships between the Shannon-Weaver index and the relative amount of average algal cell size [10], as well as the average mass of an individual in zooplankton communities [21] were identified. A relationship between the W statistic values and ∆-Shannon-Weaver is due to the fact that they are based on the same principles as the relative proportion of species in abundance and biomass in communities. Therefore, information about the size structure of communities can be obtained by different methods such as graphic (ABC– Method) that ranged species abundance, and calculated ones (Δ-Shannon-Weaver and average individual mass/cell volume).

The WESI index ranged from 0.50 to 0.67 in the three lakes, and only in the Upper Kolsay was equal to 1. Its values indicated some toxic effects on photosynthesis of algae in the Lower, Middle Kolsay and the Sary-Bulak. Heavy metal concentration in the waters of the lakes was on generally low level. Their minimum concentration (except for chromium and cadmium) was observed in the Upper Kolsay. An additional toxic factor having effect on planktonic algae can be high concentrations of iron in the three lakes. Iron is an essential element for the development of planktonic algae [58], but in high concentration has a negative impact on aquatic biota [59]. These results help us to assess the lakes ecosystems state as under early stage of toxic and nutrients impact that need for future monitoring with some traditional and advances methods [1,43,60].

The dynamics of biological indicators are linked well with the distribution of the recreational load on the lakes. The Lower Kolsay Lake, that has an automobile road leading to it, is under the greatest load. Other lakes can only be reached through the pathway, by foot or on horseback, so they are visited by a fewer tourists. This anthropogenic pressure on the lakes is growing every year. According to the official data, the number of tourists visiting the Kolsay lakes increased more than twofold from 6822 to 16000 people in the period from 2007 to 2015.

The net fishing was allowed here before the establishment of the National Park on the lakes. As a result, the near-bottom water layers of the Lower Kolsay Lake accumulated a large number of fine-mesh fishing nets that clung to the trunks of fir trees, once grown on the bottom of the valley. Fish that is captured and dies in fishing nets is the source of secondary enrichment of the water by nutrients and accelerates the process of eutrophication of Lower Kolsay Lake. Enrichment of ecosystem by nutrients is primarily manifested in the coastal shallow and more warmed-up areas of the lake. The results of future monitoring can be used as a base for the construction the regional program for the studied lakes watershed management that can to include not only hydrological and biotic data analysis but also economic and nature conservation relations [1,60].

The ongoing climate change can be taken into account in addition to the human factor. The global warming primarily influences the hydrological regime of water ecosystems in arid conditions of Kazakhstan, as well as in other regions [61]. The level of the Kolsay lakes has

fallen during recent decades: the Upper Kolsay – by  $2.5 - 3.0$  m, the Lower Kolsay – by  $0.4 - 0.5$ m. The level of Middle Kolsay Lake, that benefits from the Kolsay River and two more small rivers that originates from the western slopes, has remained almost unchanged. The hydrological regime has both indirect and direct effect on hydrocenosis and water quality through the changes of depth, nutrient concentrations, and ratio of the bottom area to water volume. Generally, the symptoms of eutrophication of aquatic ecosystems are more clearly manifested in periods of low water levels [62,63] that we have observed in the studied lakes. For the Kolsay lakes, it stays unclear what is the major cause of eutrophication – an increase of recreational load or the climate warming and the following decline of their levels. Obviously, these two factors act together and can be monitored in the Kolsay lakes placed in protected areas. That means that we can manage the water quality in these lakes only as a result of monitoring with traditional as well as modern remote sensing methods [1,60].

### **5. CONCLUSION**

The complex evaluation of the ecological state of the Kolsay high mountain lakes is set forth. According to hydrochemical variables, the lakes are classified as clean, with a slightly increased level of organic pollution in the Lower and Middle Kolsay. The ratio of species-indicators in the phytoplankton communities was reflected freshwater non-flowing alkaline waters. Regarding the ratio of species-indicators, the distribution of the functional groups (nutrition type) and indicators of size structure of phytoplankton, the level of organic pollution decreased in the high-altitude direction. The relationship between the W statistic of Clarke and the Shannon-Weaver index is shown. According to the values of phytoplankton biomass, the water of the Lower, Middle and Upper Kolsay was assessed as very clean, and of the Sary-Bulak Lake as extremely clean. Dynamics of saprobiological and dimensional parameters of phytoplanktocoenosis are linked well with the distribution of the recreational load on the lake.

The change in biomass of planktonic algae was determined to a greater extent by the dynamics of nitrogen compounds together with the<br>exhaustion of phosphorus. Analysis of of phosphorus. Analysis of<br>al and biological data is hydrochemical and biological data is demonstrated that the Kolsay mountain lakes are in the early stages of eutrophication. The eutrophication processes are most evident in the Lower Kolsay. The values of the WESI-index indicated some toxic effects on photosynthesis of algae in the Lower and Middle Kolsay and the Sary-Bulak with generally low level of heavy metals content in the water. This conclusion is confirmed by the change in the concentrations of heavy metals and iron in the water of lakes. The changes in ecosystems of the Kolsay mountain lakes can be caused by the growth of recreational load with climate change in the background.

### **ACKNOWLEDGEMENTS**

This work was partially carried out within the project G.2015 commissioned by Committee of Science, Ministry of Education and Science of the Republic of Kazakhstan "Development of methods of control of ecological state of water bodies of Kazakhstan." This work has been partly supported by the Israeli Ministry of Absorption.

#### **COMPETING INTERESTS**

Authors have declared that no competing interests exist.

### **REFERENCES**

- 1. Chang NB, Imen S, Vannah B. Remote sensing for monitoring surface water quality status and ecosystem state in relation to the nutrient cycle: A 40-year perspective. Critical Rev. Envir. Sci. Technol. 2015;45:101-166.
- 2. Busing N. Seasonality of phytoplankton as an indicator of trophic status of the large perialpine 'Lago di Garda'. Hydrobiologia. 1998;369-370:153–162.
- 3. Barinova SS, Medvedeva LA, Anissimova OV. Diversity of algal indicators in environmental assessment. Tel Aviv: Pilies Studio; 2006.
- 4. Sharov AN. Phytoplankton as an indicator in estimating long-term changes in the water quality of large lakes. Water Resources. 2008;35(6):668–673. ISSN: 0097-8078.
- 5. Kolkowitz R, Marson M. Oekologie der Tierischen Saprobien. Inter. Revue d. Ges. Hydrologie und Hydrographie. 1908;2:126– 152.
- 6. Romanenko VD, Oksiyuk OP, Zhukinsky VN, Stolberg FV, Lavrik VI. Ecological and

sanitary classification of quality of inland surface waters. Environmental impact assessment of hydraulic engineering construction on water bodies. Kiev: Naukova Dumka; 1990. Russian.

- 7. Bilous O, Barinova S, Klochenko P. Phytoplankton communities in ecological assessment of the Southern Bug River upper reaches (Ukraine). Ecohydrology and Hydrobiology. 2011;12(3):211–230. DOI: 10.2478/v10104-012-0021-3
- 8. Török P, T-Krasznai E, B-Béres V, Bácsi I, Borics G, Tóthmérész B. Functional diversity supports the biomass–diversity humped-back relationship in phytoplankton assemblages. Functional Ecology; 2016. DOI: 10.1111/1365-2435.12631
- 9. Goncharov AV. Comparison of reservoirs in the Moskva–Vazuza water system in terms of phytoplankton abundance and eutrophication degree. Water Resources. 2007;34(1):70–74. ISSN: 0097-8078.
- 10. Barinova S, Chekryzheva T. Phytoplankton dynamic and bioindication in the Kondopoga Bay, Lake Onego (Northern Russia). J. Limnol. 2014;73(2):282–297. DOI: 10.4081/jlimnol.2014.820
- 11. Berthon V, Bouchez A, Rimet F. Using diatom life-forms and ecological guilds to assess organic pollution and trophic level in rivers: a case study of rivers in southeastern France. Hydrobiologia; 2011. DOI: 10.1007/s10750-011-0786-1 (Published online 18 June 2011)
- 12. Barinova S, Klymiuk V, Lyalyuk N. Ecology of phytoplankton in the regional landscape park "Slavyansky Resort", Ukraine. Applied Ecology and Environmental Research. 2015;13(2):449–464. DOI: 10.15666/aeer/1302\_449464
- 13. Klymiuk V, Barinova S. Phytoplankton Cell Size in Saline Lakes. Research Journal of Pharmaceutical. Biological and Chemical Sciences. 2016;7(1):1077–1085.
- 14. Warwick RM. A new method for detecting pollution effects on marine macrobenthic communities. Mar. Biol. 1986;92(4):557– 562.
- 15. DelValls TA, Conradi M, Garcia-Adiego E, Forja JM, Gómez-Parra A. Analysis of macrobenthic community structure in relation to different environmental sources of contamination in two littoral ecosystems from the Gulf of Cádiz (SW Spain). Hydrobiologia. 1998;385:59–70.
- 16. Rakocinski CF, Brown SS, Gaston GR, Heard RW, Walker WW, Summers JK. Species-abundance-biomass responses by estuarine macrobenthos to sediment chemical contamination. Journal of Aquatic Ecosystem Stress and Recovery. 2000;7: 201–214.
- 17. Harkantra SN, Rodrigues NR. Numerical analyses of soft bottom macroinvertebrates to diagnose the pollution in tropical coastal waters. Environmental Monitoring and Assessment. 2004;93:251–275.
- 18. Krupa EG, Grishaeva OV. The structure of the species dominance in macrozoobenthos in the Small Aral Sea as an indicator of changes in water salinity. In: Rumyantsev VA, editor. Bioindication in the monitoring the freshwater ecosystems. St Petersburg: "Lubavitch"; 2011. Russian.
- 19. Krupa EG. The structure of species dominance in the plankton of the Caspian Sea of Kazakhstan sector. In: Malkovskiy IM, editor. Several aspects of hydroecological problems of Kazakhstan. Kaganat: Almaty; 2011. Russian.
- 20. Krupa EG. Rank distributions of species in the zooplankton communities in Kazakhstan reservoirs as an indicator of their ecological condition. In: Rumyantsev VA, editor. Bioindication in the monitoring of freshwater ecosystems. St Petersburg: Lubavitch"; 2011. Russian.
- 21. Krupa EG. Zooplankton of limnetic and lothic ecosystems of Kazakhstan. Structure, patterns of formation.<br>Saarbrucken: Palmarium Academic Saarbrucken: Publishing; 2012. Russian.
- 22. Kurmangaliyeva SG. Seasonal dynamics of zooplankton in the Lower Kolsay Lakes. Biological Sciences. 1974;7:87–91.
- 23. Smirnova DA. State of zooplankton of the Middle and Lower Kolsay Lakes (Chilik river watershed) during the beginning of their recreational use. Kazakn National University Bulletin, Biology Series. 2000;4: 54–60. Russian
- 24. Tsugeki NK, Agusa T, Ueda Sh, Kuwae M, Oda H. et al. Eutrophication of mountain lakes in Japan due to increasing deposition of anthropogenically produced dust. Ecol Res. 2012;27:1041–1052. DOI: 10.1007/s11284-012-0984-y
- 25. Semenova AD, editor. Guidelines for chemical analysis of surface water. Leningrad: Gidrometeoizdat; 1977. Russian.
- 26. Fomin GS. Water. Control of chemical, bacterial and radiation safety according to international standards. Moscow: NGO "Alternative"; 1995. Russian.
- 27. Kimstach VA. Classification of surface water quality in the European Economic Community countries. St Petersburg: Gidrometeoizdat; 1993. Russian.
- 28. Bespamyatnov GP, Krotov YA. Maximum allowable concentrations of chemicals in the environment. Leningrad: Chemistry; 1985. Russian.
- 29. Kiselev IA. Methods to study plankton. In the book: The life of fresh waters in the USSR. Moscow, Leningrad: Science Academy of the USSR. 1956;4. Russian.
- 30. Hollerbach MM, Kosinskaya EK, Polyansky VI. Determinant for freshwater algae of the USSR. The blue-green algae. Moscow: Soviet Science.1953;2. Russian.
- 31. Matvienko AM. Determinant for freshwater algae of the USSR. Golden algae. Moscow: Soviet Science. 1954;3. Russian
- 32. Moshkova NA, Hollerbah MM. Determinant for freshwater algae of the USSR. Green algae. Class ulotrichous. Order ulotrichous. Moscow: Soviet Science. 1986;10:1. Russian.
- 33. Palamar-Mordvintseva GM. Determinant to freshwater algae of the USSR. Green algae. Class conjugates. Order desmid (2). Moscow: Soviet Science. 1982;11:2. Russian.
- 34. Popova TG. Determinant to freshwater algae of the USSR. Euglena algae. Moscow: Soviet Science. 1955;7. Russian.
- 35. Zabelina MM, Kiselev IA, Proshkina-Lavrenko AI, Sheshukova VS. Determinant for freshwater algae USSR. Diatoms. Moscow: Soviet Science. 1951;4. Russian.
- 36. Magurran AE. Ecological diversity and its Measurement. Moscow: World; 1998. Russian.
- 37. Clarke KR. Comparison of dominance curves. J. Exp. Mar. Biol. Ecol. 1990;138: 143–157.
- 38. Heywood V. Modern approaches to floristics and their impact on the region of SW Asia. Turk. J. Bot. 2004;28:7–16.
- 39. Ter Braak CJF, Šmilauer P. CANOCO reference manual and CanoDraw for windows user's guide: Software for canonical community ordination (version 4.5). Ithaca: Microcomputer Power Press; 2002.
- 40. Ter Braak CJF. The analysis of vegetationenvironment relationships by canonical

correspondence analysis. Vegetatio. 1987; 69:69–77.

- 41. Ter Braak CJF. Interpreting canonical correlation analysis through biplots of structural correlations and weights. Psychometrica. 1990;55:519–531.
- 42. Novakovsky AB. Abilities and base principles of program module "GRAPHS". Scientific Reports of Komi Scientific Center, Ural Division of the Russian Academy of Sciences. 2004;27:1–28.
- 43. Barinova S. Algal diversity dynamics, ecological assessment, and monitoring in the river ecosystems of the eastern Mediterranean. New York: Nova Science Publishers; 2011.
- 44. Hustedt F. Systematische und Ökologische Untersuchungen über die Diatomeenflora von Java, Bali und Sumatra. Arch. Hydrobiol. Suppl. 1938– 1939;15:131–177, 393–506, 638–790; 16: 1–155, 274–394.
- 45. Hustedt F. Die diatomeenflora des Flußsystems der weser im gebiet der hansestadt bremen. Abhandl. Naturwis Ver Bremen. 1957;34:181–440.
- 46. Watanabe T, Asai K, Houki A. Numerical estimation to organic pollution of flowing water by using the epilithic diatom assemblage – Diatom Assemblage Index (DAIpo). Science Total Environment. 1986; 55:209–218.
- 47. Van Dam H, Martens A, Sinkeldam J. A coded checklist and ecological indicator values of freshwater diatoms from the Netherlands. Netherlands J. Aquatic Ecol. 1994;28(1):117–133.
- 48. Slivinsky GG, Krupa EG, Lopatin OE, Mamilov NS, Prikhodko DE. Ecological and toxicological characteristics and condition of the aquatic fauna of the Tekesskiy reservoir. Kazakn National University Bulletin. Ecology Series. 2010;1:79–88. Russian.
- 49. Slivinsky GG, Krupa EG, Mamilov NS, Akhmetov EM. Environmental condition and the state of hydrofauna of the rivers Kegen and Sharyn. In: Alghanov EM, editor. Proceedings of Charyn State National Park. Kaganat, Almaty. 2013;1. Russian.
- 50. Ahmedenov KM, Koshim AG. The geoecological characteristics of spring valleys of West Kazakhstan region. Herald Kazakh National University. Geographical series. 2014;2(39):18-23. Russiian.
- 51. Xu H, Min G, Choi J, Al-Rasheid KAS, Lin X, Zhu M. Temporal dynamics of phytoplankton communities in a semienclosed mariculture pond and their responses to environmental factors. Chinese Journal of Oceanology and Limnology. 2010;28(2):295–303. DOI: 10.1007/s00343-010-9257-1
- 52. Tolotti M, Manca M, Angeli N, Morabito G, Thaler B, Rott E, Stuchlik E. Temperature modulated effects of nutrients on phytoplankton changes in a mountain lake. Hydrobiologia. 2012;698:61–75. DOI: 10.1007/s10750-012-1146-5
- 53. Mingli Y, Cuixia ZH, Zengjie J, Shujin G, Jun S. Seasonal variations in phytoplankton community structure in the Sanggou, Ailian, and Lidao Bays. J. Ocean Univ. China (Oceanic and Coastal Sea Research). 2014;13(6):1012–1024. DOI: 10.1007/s11802-014-2305-2
- 54. Odum E. Ecology. Moscow: World; 1981. Russian.
- 55. Kryuchkova NM. The structure of the zooplankton communities in waters of different types. In: Alimov AF, editor. Production and hydrobiological studies of aquatic ecosystems. Leningrad: Science; 1987. Russian.
- 56. Andronicova IN. Structural and functional organization of zooplankton in lake ecosystems of different trophic types. St Petersburg: Science; 1996. Russian.
- 57. Kókai Z, Bácsi I, Török P, Buczkó K, T-Krasznai E, Balogh C, et al. Halophilic diatom taxa are sensitive indicators of even short term changes in lowland lotic systems. Acta Bot. Croat. 2015;74(2): 287–302. DOI: 10.1515/botcro-2015-0025
- 58. Vrede T, Tranvik LJ. Iron constraints on Planktonic primary production in Oligotrophic Lakes. Ecosystems. 2006;9: 1094–1105. DOI: 10.1007/s10021-006-0167-1
- 59. Bakker ES, Donk EVI, Anne K. Lake restoration by in-lake iron addition: A synopsis of iron impact on aquatic organisms and shallow lake ecosystems. Aquat. Ecol. 2016;50:121–135. DOI: 10.1007/s10452-015-9552-1
- 60. Chang NB, Wen CG, Wu SL. Optimal management of environmental and land resources in a reservoir watershed by multi-objective programming. J. Envir. Manag. 1995;44:145-161.
- 61. Jeppesen E, Brucet S, Naselli-Flores L, Papastergiadou E, Stefanidis K, Noges T, et al. Ecological impacts of global warming and water abstraction on lakes and reservoirs due to changes in water level and related changes in salinity. Hydrobiologia. 2015;750:201–227. DOI: 10.1007/s10750-014-2169-x
- 62. ZhengJian Y, DeFu L, DaoBin J, ShangBin X. Influence of the impounding process of the Three Gorges Reservoir up to water level 172.5 m on water eutrophication in the Xiangxi Bay. Science China. Technological Sciences. 2010;53(4):1114– 1125.

DOI: 10.1007/s11431-009-0387-7

63. Chuai X, Chen X, Yang L, Zeng J, Miao A, Zhao H. Effects of climatic changes and anthropogenic activities on lake eutrophication in different ecoregions. Int. J. Environ. Sci. Technol. 2012;(9):503– 514.

DOI: 10.1007/s13762-012-0066-2

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